

# 1 Prioritised substance group: Bisphenols

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## 1.1 Background Information

### 1.1.1 Hazardous properties

#### Bisphenol A

There is a large amount of literature on the toxicity of bisphenol A including at low doses [reviewed in WHO and UNEP (2012), Gore et al. (2015), Vandenberg (2014), and EFSA Journal (2015)].

Studies have indicated that it could be associated with increased risk for:

- ▶ Fetal development: miscarriages, decreased birth weight at term,
- ▶ Reproductive and sexual dysfunctions,
- ▶ Breast and prostate cancer or at least significant breast tissue remodelling. Studies have indicated that those effects were associated with gestational and neonatal exposure [Seachrist et al. (2016)].
- ▶ Altered immune system activity,
- ▶ Obesity and metabolic dysfunctions and diabetes in adults,
- ▶ Cardiovascular disease in adults
- ▶ Cognitive and behavioural development in young children.

Despite the wealth of studies, there are still controversies concerning the toxic effects of BPA. Those are related to some lack of reproducibility of the experimental studies possibly due to different designs as well as on issues related to the analytical procedures used for BPA assays. Several studies (both experimental and human) have focused on perinatal exposure using different doses including low doses and monitoring a variety of outcomes [FitzGerald and Wilks (2014)]. In human there are several cohort studies associating perinatal exposure and child development. In addition, there are cross-sectional studies where associations were found between BPA exposure and metabolic and cardiovascular diseases. The latter studies have established association but cannot reveal a causal link between BPA and a toxic outcome. In conclusion, there is a real concern that BPA exposure could be linked to a variety of health outcomes in human, with different level of evidence depending on the outcome and the exposure period. Other Bisphenols, notably many BPA substitutes structurally similar to BPA, have been less studied although data suggest they are also oestrogenic, and likely to induce similar health effects [Rochester and Bolden (2015), Shalenie et al. (2020)].

BPA elicits a variety of endocrine disrupting effects targeting steroid hormones as well as thyroid hormones. Several studies have explored the mechanisms of endocrine disruption. Initial studies have indicated an interaction with the nuclear ER alpha oestrogen receptor with a relatively low affinity. Further studies have indicated an interaction with other receptors such ERbeta, ERRgamma and GPR32. An unresolved question is which of those receptors is involved in the low dose fetal effects of BPA.

Because of the controversies on BPA toxicity, a collaborative project called CLARITY-BPA was carried out in the US involving both regulatory agencies and academic laboratories [National Toxicology Program Research report 9 (2018)]. Despite the fact that animal treatment was centralised, different outcomes and conclusions were reached by different groups. The core studies run by the FDA found little consistent evidence for toxicity and for non-monotonic dose-response curves when traditional outcomes were examined [Camacho et al. (2019)]. Studies done in academic laboratories found evidence for low-dose effects and non-monotonic dose-response [Prins et al. (2019)]. As an illustration, a recent scientific study, undertaken as part of the CLARITY-BPA project, developed a quantitative assessment of the effects of bisphenol A (BPA) exposure on mammary gland development and found a consistent pattern of non-monotonic dose response relationships on a set of over 90 measurements. This demonstrates a causal relationship between exposure to BPA and the health effects observed [Montévil et al. (2020)]. The reasons for these discrepancies are unclear and they could be related to a non-optimal study design, different health outcomes, different analysis of the data and different interpretation of some data.

### **Bisphenol S, F and others**

Recent studies on BPS toxicity are published. Regarding BPS toxicity on reproduction in humans, maternal prenatal urinary BPS concentrations were consistently associated, but not significantly, with various markers of fetal growth [Ferguson et al. (2018)].

Urinary BPS was correlated with increased gestational age and increased risk of late term birth for girls [Wan et al. (2018)], and with preterm birth [Aung et al. (2019)].

In a EU cohort, no association of bisphenol analogues including BPF with fecundability was reported, but total bisphenols (including 4,4-BPF, BPS, BPB, BPP, BPAF, BPAP, or BPZ) was associated with a longer time to pregnancy in women with inadequate folic acid supplement use [Philips et al. (2018)].

BPF, BPS, BPAF, along with Bisphenol Z (BPZ), Bisphenol E (BPE) and Bisphenol B (BPB) are suspected to be endocrine disrupting chemicals which are oestrogenic [Mesnage et al. (2017)]. In human studies investigating health effects including endocrine effects of BPS and other bisphenols [see review by Pelch et al. (2019)], conflicting results have been reported for an association with obesity, diabetes, fasting blood glucose or insulin resistance for BPA analogs including BPS and BPF. Regarding effects on thyroid, BPS was associated with a suggestive increase in TSH, as well as a decrease in free T4 [Aker et al. (2019)].

In vivo scientific evidences were also released recently on BPS and organs or systems such as the mammary gland, female reproductive system, the male reproductive system and on metabolism and obesity. A developmental toxicity study on BPS according to OECD guideline 414 in pregnant rats did not reveal any reproductive, developmental or teratogenic effects (ECHA Dissemination, 2018). However, different academic papers studied the effects of BPS on male reproductive function and suggest a coherent picture of the alterations in spermatogenesis after BPS exposure in both rat and mice [Shi et al. (2018), Horan et al. (2018), Shi et al. (2017), Ullah et al. (2018), Shi et al. (2019)].

Among evidenced effects reported on female reproductive study in animals, are alteration of the pattern of oocyte maturation/meiosis and/or folliculogenesis, and, in a lesser extent, the timing of puberty. [Shi et al. (2017 and 2019), Nourian et al. (2017), Ijaz et al. (2019), Ahsan et al (2018), Horan et al. (2018), Nevoral et al. (2018)].

BPB, BPE, BPF display anti-androgenic activities in some settings [Rosenmai et al. (2014)]. Moreover a study on BPS and BPAF exposure showed that it can modify the histology of zebra fish testes and ovaries and influence homeostasis of testosterone and oestradiol, and parental exposure to environmentally relevant concentration of BPAF results in delayed hatching of the

offspring [Shi et al. (2015)]. BPS and BPF induce proliferation and migration of breast cancer cells via the oestrogen receptor dependent pathway in vitro [Kim et al. (2017)].

A new text-mining tool was developed to explore the literature and attempt link bisphenols to adverse outcome pathways (AOP-helpFinder). Using this tool as well as systems biology approaches, it was found that BPS could be linked to pathways leading to obesity [Carvaillo et al. (2019)] and BPF to AOP networks leading to thyroid cancer [Rugard et al.(2020)]. These data and others suggest that the safety of BPA substituents is not clear at this stage.

### 1.1.2 Exposure characteristics

BPA is used in certain plastics, epoxy resins and thermal papers and is among the highest volume of chemicals produced world-wide. There is evidence that contamination can occur through different routes, including food, water, air and skin (particularly in occupational exposure of cashiers). BPA has a relatively short half-life (hours); it is conjugated and believed to be inactive in that form, but there is concern that it may be locally deconjugated at the tissue level. There is a clear advantage in measuring free and conjugated forms both to address the possibility of external contamination during the assay and to better assess the active form of the substance.

There is solid evidence that a large majority of the human population is exposed to BPA. Many biomonitoring studies are available for bisphenol A (BPA) but the majority of the studies have a single measurement of exposure. These studies are useful in estimating the exposure to BPA in a particular population and follow time trends but not for risk assessment. Studies with multiple biological samples (usually pregnancy cohorts) have shown that BPA has poor Intraclass Correlation Coefficient (ICC) and therefore a single biological measurement can cause exposure misclassification. Further, there is a lack of consensus on how to deal with multiple samples in estimating the correct exposure. In addition, not all countries in Europe have biomonitoring data available on BPA. In DEMOCOPHES<sup>1</sup>, seventeen European countries participated, but BPA was added for a group of only 6 countries. BPA is analysed in very few European birth cohorts in Germany, Norway, Spain and France [Casas et al. (2013)].

Bisphenol F (BPF), Bisphenol S (BPS), and Bisphenol AF (BPAF) are among the main substitutes of BPA [Chen et al. (2016); Gao et al. (2020); Yang Y et al. (2019)]. Studies in food revealed that BPS and some other bisphenols can be detected besides BPA in a large number of foodstuffs at low concentrations [Vinas et al. (2010); Liao and Kannan, (2013, 2014)].

As part of the national biomonitoring program, the Esteban cross-sectional study has measured, for the first time in the continental French population, the levels of impregnation with bisphenols A, S and F. The measurement of urinary concentrations of bisphenols was based on a subsample of 500 children and 900 adults, aged 6 to 74, included in the study between April 2014 and March 2016. Bisphenols A, S and F were detected in almost all samples; the geometric mean in BPA was 2.25 and 2.69 µg / g creatinine, respectively, in children and adults; equivalent to 0.44 and 0.53 µg / g creatinine for bisphenol S (BPS), and 0.26 and 0.31 µg / g creatinine for bisphenol F (BPF). Impregnation with bisphenols was higher in children than in adults. The results obtained were close to those observed in North American countries [Santé Publique France (2019)].

### 1.1.3 Policy relevance

Regulatory measures have been taken at the EU level while additional measures have been taken in certain countries. In the EU, bisphenol A is regulated under REACH (1907/2006/EC). EU law regulates BPA in plastic materials and articles intended to come into contact with food

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<sup>1</sup> Demonstration Of A Study To Coordinate And Perform Human Biomonitoring On A European Scale – DEMOCOPHES (2010) <http://www.eu-hbm.info/democophes>

[Commission Regulation (EU) No 10/2011], and since 2011 BPA has been banned from infant feeding bottles across Europe [Commission Directive 2011/8/EU]. In 2018, the EU further restricts the use of bisphenol A in certain food-contact materials. A specific migration limit (SML) for BPA in varnishes and coating has been introduced and the SML for BPA in the Plastics Regulation has been revised. [Commission Regulation (EU) 2018/213]. Additional measures have been taken in several countries. For example, France banned BPA in all food contact materials [French Law No 2012-1442], other countries like Denmark, Belgium and Sweden, banned it in those materials intended for children under 3.

Since 2017 BPA is on the Candidate List of substances of very high concern for Authorisation (SVHC candidates) as it is classified toxic for reproduction. France has prepared a dossier for the identification of BPA as a human ED-SVHC substance, and Germany for the identification as an environmental ED-SVHC substance. In June 2017, ECHA identified BPA as a substance of very high concern (SVHC) due to alleged endocrine disrupting (ED) effects for human health and the environment [ECHA (2017)]. In October 2019, ECHA prioritised BPA for toughest EU restrictions by proposing its use should be subject to prior authorisation.

There are also controversies between different agencies concerning the most protective Total Daily Intake (TDI). Furthermore, BPA is also present in thermal papers and exposure of cashiers has been assessed and led to a proposal for restriction and substitution. Different committees of ECHA have analysed the benefits and costs of restrictions and sent their conclusion to the European Commission. BPA is restricted in the EU in thermal paper since 2016. The ban has taken effect in January 2020, giving companies time to phase it out and find a safer alternative. BPA is being primarily replaced by BPS in thermal paper, however it is likely not a safer alternative. In Switzerland, BPS is banned from thermal paper (at a concentration equal to or greater than 0.02% by weight) as well, as of June 2020.

Currently in the EU, there is a limit on the amount of BPA that is allowed to leach out of toys for children up to the age of three and in any toys that are intended to be placed in a child's mouth. The migration limit has been decreased to **0.04 mg/l in toys [Commission Directive (EU) 2017/898]**.

BPA regulation is actively debated across the world. BPS and BPF are the major BPA substituents with distinct industrial applications. Much less is known about their putative toxicity and their presence in human matrices, although initial studies have indicated that they may display toxic effects that are similar to BPA [Rochester and Bolden (2015), Auerbach et al. (2016)]. ECHA has started the process for a harmonised classification and labelling on reproductive toxicity for BPS<sup>2</sup>. The consultation period was closed on February 2020. In addition, a similar consultation recently finished for BPAF<sup>3</sup>. Other bisphenol compounds are also manufactured and little is known about their toxicity and diffusion at this stage.

#### **1.1.4 Technical aspects**

Although BPA (and to a much lesser extent BPS and BPF) have been assayed in several Human Biomonitoring studies there is a need to harmonise procedures for sample handling, storage and analytical methodologies. However, assays for conjugated and free substances should also be harmonised. The same holds true for other bisphenols.

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<sup>2</sup> ECHA 2019; [BPS](#) - CLH report Proposal for Harmonised Classification and Labelling Based on Regulation (EC) No 1272/2008 (CLP Regulation), Annex VI, Part 2

<sup>3</sup> ECHA 2019; [BPAF](#) - CLH report Proposal for Harmonised Classification and Labelling Based on Regulation (EC) No 1272/2008 (CLP Regulation), Annex VI, Part 2

Furthermore, external contamination during sample collection, handling and analysis is an important criteria during the evaluation of studies to be considered both for assigning reference values (HBM values) and risk assessment. For BPF and BPS, there are few biomonitoring studies available (see below) but there is a lack of literature for other bisphenols [Chen et al. (2016)].

### **1.1.5 Societal concern**

In several countries and probably world-wide, BPA has been considered as the typical endocrine disruptor. In many cases, the societal concern towards EDCs is highly connected to the bisphenol case and to the campaigns to regulate BPA. Therefore there is a lot of expectations in this field. It is important to fill the gaps and to attempt to address the uncertainties, because the bisphenol case appears to be emblematic of the EDC. Whatever we achieve with bisphenols will actually be useful for all EDCs and for the role of public authorities in protecting pregnant women and the next generations.

## 1.2 Categorisation of Substances

Table 1-1: Substances included in the substance group, listed according to availability of toxicology and human biomarker data, in category A, B, C, D, E substances (see general introduction)

Category	Abbreviation/ Acronym	Systematic name	CAS No.	Regulation
A	BPA	4,4'-isopropylidenediphenol	80-05-7	REACH Annex V; Annex XVII, Entry 66 PACT list entry 13/04/2017: Hazard assessment. Scope: ED. <u>OSH Legislation</u> : Consumer uses, Article service life, Widespread uses by professional workers, Formulation or re-packing, Uses at industrial sites, Manufacture, Signs at work, CAD, Young workers, Pregnant or breastfeeding workers. <u>Environmental legislation</u> : Classified Seveso Waste Directive Annex III <u>Professional and consumer legislation</u> : Cosmetics (EC) No 1223/2009 Annex II; Toy safety Directive Appendix C
C	BPS	4,4'-sulphonyldiphenol	80-09-1	CoRAP list PACT list entry 01/04/2015: Hazard assessment. Scope: ED. <u>OSH Legislation</u> : Article service life, Formulation or re-packing, Uses at industrial sites, Manufacture.
C	BPF	4,4'-methylenediphenol	620-92-8	REACH Annex III Inventory PACT list entry 01/10/2015: RMOA. Scope: ED. Under development <u>OSH Legislation</u> : CAD
C	BPB	4,4'-(1-methylpropylidene) bisphenol	77-40-7	REACH Annex III Inventory PACT list entry 07/03/2017: RMOA. Scope: ED. Under development <u>OSH Legislation</u> : CAD
C	BPAF	4,4'-[2,2,2-trifluoro-1-(trifluoromethyl)ethylidene] diphenol	1478-61-1	REACH Annex III Inventory PACT list entry 01/10/2015: RMOA. Scope: ED. Under development <u>OSH Legislation</u> : CAD <u>Environmental legislation</u> : Classified Seveso

Category	Abbreviation/ Acronym	Systematic name	CAS No.	Regulation
C	BPAP	4,4'-(1-Phenylethylidene) bisphenol	1571-75-1	REACH Annex III Inventory <u>OSH Legislation</u> : CAD <u>Environmental legislation</u> : Classified Seveso Waste Directive Annex III
C	BPBP	2,2-bis(2-hydroxy-5-biphenyl)propane	24038-68-4	-
C	BPC	4,4'-isopropylidenedi-o-cresol	79-97-0	REACH Annex III Inventory
C	BPCI2	4,4'-(dichlorovinylidene) diphenol	14868-03-2	REACH Annex III Inventory <u>OSH Legislation</u> : CAD
C	BPE	4,4'-Ethylidenebisphenol	2081-08-5	-
C	BPPH	4,4'-Dihydroxytetraphenylmethane	1844-01-5	-
C	BPM	4,4'-(1,3-phenylene-bis(1-methylethylidene))bis-phenol	13595-25-0	CoRAP list PACT list entry 02/02/2017: Hazard assessment. Not ED. <u>OSH Legislation</u> : Article service life, Uses at industrial sites, Manufacture, Signs at work, CAD, Pregnant or breastfeeding workers. <u>Environmental legislation</u> : Classified Seveso Waste Directive Annex III
C	BPP	4,4'-(1,4-Phenylene diisopropylidene)bisphenol	2167-51-3	REACH Annex III Inventory <u>OSH Legislation</u> : CAD
C	BIS2	Bis(2-hydroxyphenyl) methane	2467-09-9	-
C	DHDPE	p,p'-oxybisphenol	1965-09-9	REACH Pre-Registration process <u>OSH Legislation</u> : CAD, Young workers. Waste Directive Annex III
C	BPFL	9,9-Bis(4-hydroxyphenyl) fluorene	3236-71-3	REACH Registration Dossier <u>OSH Legislation</u> : CAD <u>Environmental legislation</u> : Classified Seveso Waste Directive Annex III

Category	Abbreviation/ Acronym	Systematic name	CAS No.	Regulation
C	BPZ	4,4'-cyclohexylidene bisphenol	843-55-0	REACH Annex III Inventory <u>OSH Legislation</u> : CAD, Pregnant or breastfeeding workers. <u>Environmental legislation</u> : Classified Seveso
C	BP4,4'	Biphenyl-4,4'-diol	92-88-6	REACH Registration Dossier <u>OSH Legislation</u> : Widespread uses by professional workers, Uses at industrial sites, CAD, Young workers. <u>Environmental legislation</u> : Classified Seveso Waste Directive Annex III <u>Professional and consumer legislation</u> : Plastic contact with food (EU) No 10/2011 Annex I

### 1.3 Policy-related questions

There are several critical questions concerning bisphenols that need to be resolved. The first is whether different regulations in different countries lead to different internal exposure values and whether the increasingly frequent use of substituents has led to increased exposure and to the presence of mixtures of bisphenols in humans. The second is identify safety values taking into consideration the accumulating knowledge on Bisphenol toxicity particularly at low doses. A third question is whether substitutes are safer than BPA considering their hazardous properties and current and expected exposure to those compounds.

Specific policy-related questions are:

1. What is the current exposure of the EU population to BPA, BPS and BPF?
2. Do different regulatory controls across the EU concerning in particular BPA lead to different exposures?
3. Are bisphenols (BPA and substitutes) exposure levels of concern for health?
4. Is occupational exposure of cashiers (BPA and substitutes) a health concern?
5. What is the toxicity of BPA substitutes?
6. Are health risks age and gender dependent?
7. Can we find evidence for low-dose effects within mixtures?
8. How can HBM feed into assessment of the Tolerable Daily Intake (TDI) for BPA, as set by the European Food Safety Authority (EFSA)?
9. Is it important to eliminate legacy BPA from material cycles (i.e. waste till receipt rolls) when implementing a circular economy in order to protect human health?

## 1.4 Research Activities to be undertaken

Table 1-2: Listing of research activities to be carried out to answer the policy questions summed up in 1.3

Policy question	Substance	Available knowledge	Knowledge gaps and activities needed
1. What is the current exposure of the EU population to BPA, BPS and BPF?	BPA	Different types of studies could be considered to address the current exposure of the EU population to BPA: well characterised samples such as Cophes/Democophes, high quality national studies, studies including several samples per individual to account for intra-individual variability, studies with available or planned health outcomes. In the annex to table 2 following EU countries are listed to have HBM data on BPA: De, Be, Fr, At, Cz, Se, El and DEMOCOPHES countries. Age groups for which data are available differ among the countries.	Inventorying available HBM data in EU ->WP7 Use already available biomonitoring data in Europe to: 1. Find out which countries lack HBM data 2. Evaluate the quality of the available data such that design of future biomonitoring studies can be improved accordingly, calculate exposure levels. -> WP8, WP9, WP10 To use already available biomonitoring data in Europe to define the minimum number of samples required per individual to estimate the correct exposure to BPA. Ensure that aggregated data are available in IPCHEM and calculate reference value for BPA base on existing data-> WP10; Q:1,2 To compare single samples vs multiple samples for exposure assessment -> WP9, WP13 To identify exposure pathways for BPA and its toxicokinetic characteristics. -> WP12
2. Do different regulatory controls across the EU concerning in particular BPA lead to different exposures?	BPA (Cat. A)	On 16 June 2017 ECHA <sup>4</sup> classified BPA as an endocrine disruptor and a substance of high concern. Specific countries such as France, Denmark and Sweden have already stricter bans in place. ECHA <sup>5</sup> , has recently updated the BPA entry to reflect an additional reason for inclusion due to its endocrine disrupting properties causing adverse effects to the environment.	Find out whether there are HBM data or suitable samples available before and after the ban in Fr, Se, Dk? If not, design an appropriate study to analyse samples time trends-> WP7, WP8 Increase knowledge on time trends, geographical comparisons, exposure determinants of existing HBM data and new HBM data (WP9/WP10; Q: 1, 2) Complete the analysis of collected samples from the alignment study and initiate data exploitation and analysis. HBM data for BPA, BPS and BPF (WP8/WP9/WP10; Q: 1, 2)

<sup>4</sup> <https://echa.europa.eu/-/msc-unanimously-agrees-that-bisphenol-a-is-an-endocrine-disruptor>

<sup>5</sup> <https://echa.europa.eu/-/seven-new-substances-added-to-the-candidate-list-entry-for-bisphenol-a-updated-to-reflect-its-endocrine-disrupting-properties-for-the-environment>

Policy question	Substance	Available knowledge	Knowledge gaps and activities needed
<p>3.</p> <p>Are bisphenols exposure levels of concern for health?</p>	<p>BPA (Cat. A)</p> <p>BPS, BPF (Cat. C)</p>	<p>KEMI<sup>6</sup> has identified a total of over 200 chemical substances that have a chemical structure similar to Bisphenol A and which may be found on the European market. According to results from data simulations, 37 of these substances may have endocrine disrupting properties like those identified for Bisphenol A.</p> <p>Only available in vitro/in vivo experimental settings in which BPA AOP have already been explored will be used to assess the effects of BPS and BPF.</p>	<p>To determine whether current or expected exposure levels of BPS and BPF are of concern for health in the general population and at the workplace?</p> <p>Derive health based HBM guidance values for the general population and for workers (WP5 / T5.2) and to identify the relationship to the environment and workplace. What is the evidence for low-dose effects? Analyse data from longitudinal cohort studies -&gt; WP13</p> <p>Do BPS and BPF act on the same AOPs as BPA? -&gt; WP13</p> <p>Targeted assessment of toxic effects of BPS/BPF as compared to BPA. Targets priority will be given to cancer, reproductive, hormonal, metabolic, immune and neurological effects. The linkage with effect biomarkers could be explored in human samples as well as mixture effects. -&gt; WP13 / T13.1, WP14 / T14.3 and WP15 / T15.3.</p> <p>Urgently harmonise procedures for sample handling, storage and analytical methodologies for BPA, BPS and BPF to minimise external contamination. Encourage European countries to participate in inter-laboratory comparisons. Optimise the analytical method for BPS and BPF. Establish the definitive database with those labs which successfully achieved quality criteria for bisphenols assays (BPA, BPS and BPF) -&gt; WP9</p> <p>To identify existing analytical methods allowing to monitor in human matrices BPA, BPS, BPF and possibly other BPs, as well as the necessary gaps to be fulfilled in terms of method development and validation. -&gt; WP9, WP16</p> <p>To identify exposure pathways for BPS, BPF (possibly other BPs) and their toxicokinetic characteristics. -&gt; WP12</p> <p>To identify effect biomarkers associated to bisphenol exposure and to determine whether those effect markers are common to all bisphenol compounds. -&gt; WP14</p>

<sup>6</sup> <https://www.kemi.se/en/news-from-the-swedish-chemicals-agency/2017/new-report-37-bisphenols-may-be-endocrine-disruptors/>

Policy question	Substance	Available knowledge	Knowledge gaps and activities needed
4. Is occupational exposure of cashiers a health concern?	BPA (Cat. A)	BPA is restricted in the EU in thermal paper since 2016 <sup>7</sup> . The ban takes effect in 2020. A directive <sup>8</sup> revised the OEL (Occupational Exposure Limit values) for BPA of 2mg/m <sup>3</sup> TWA in occupational settings.	To find out whether BPA occupational exposure of cashiers is a health concern. To feed into the Commission decision on whether to ban BPA in till receipts, as recommended by ECHA's Committee for Socio-Economic Analysis (SEAC).-> WP8 Finalise external and internal modelling predictions for bisphenol A, F and S (WP12; Q: 2, 3, 4, 8)
5. What is the toxicity of BPA substitutes?	BPS, BPF (Cat. C)	BPS and BPF are the primary substitutes of BPA. Some countries have started the process to restrict BPS because of its toxicity profile. DG Grow and DG Santé recommend to monitor BPA as well as BPS and BPF, the most prominent substituents. Recent findings in the US <sup>9</sup> have shown that some people's exposure to BPF can meanwhile be higher than to BPA. This evidence should also be explored in Europe particularly in women of child-bearing age.	What is the toxicity of BPA substitutes? To identify effect biomarkers associated to bisphenol exposure and to determine whether those effect markers are common to all bisphenol compounds. -> WP14 Finalise the studies on Adverse Outcome Pathways (AOPs) for BPA, BPF, BPS, identify gaps and apply new approaches linking those chemicals to AOPs such as quantitative AOPs (WP13; Q: 3, 5, 6, 7) Exploit cohorts based on existing data and summarise links between bisphenols and health outcomes. Focus on male fertility endpoints and endocrine disruptive effects. (WP13; Q: 3, 5, 6, 7)
6. Are health risks age and gender dependent?	BPA BPS, BPF (Cat. C) Other Bisphenols (Cat. C)	Most regulation and recommendation tend to focus on pregnant women and infants.	To determine age and gender specific health effects of BPA. -> WP9, WP10, WP13 To derive health based HBM guidance values and perform risk assessments for different age groups and sex. -> WP5 / T 5.2 & 5.3 To measure hormonal levels in new-borns exposed transplacentally and pubertal children due to environmental exposure to BPs. -> WP14

<sup>7</sup> <http://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32016R2235&from=FR>

<sup>8</sup> <http://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32017L0164&from=FR>

<sup>9</sup> <https://silentspring.org/blog/results-our-biomonitoring-study-are>

Policy question	Substance	Available knowledge	Knowledge gaps and activities needed
<p>7.</p> <p>Can we find evidence for low-dose effects within mixtures?</p>			<p>To determine the effect of combined exposures to substance mixtures within the bisphenol family and with other families and whether this could impact health guidance (in food contamination, cosmetics, other plasticizers, etc.) -&gt; WP15</p> <p>To identify effect biomarkers associated to bisphenol exposure and to determine whether those effect markers are common to all bisphenol compounds. Assay selected effect markers in the alignment studies and link them to BP levels -&gt; WP14</p> <p>To develop indicators for combined exposures. -&gt; WP5 / T 5.4</p>
<p>8.</p> <p>How can HBM feed into assessment of the Tolerable Daily Intake (TDI) for BPA, as set by the European Food Safety Authority (EFSA)?</p>	<p>BPA (Cat. A) BPS, BPF Other Bisphenols (Cat. C)</p>	<p>In its 2015 re-evaluation of BPA exposure and toxicity, EFSA (European Food Safety Authority) used a more refined methodology than before supported by new data to review the Tolerable Daily Intake (TDI) for BPA from 50 to 4 µg/kg bw/day.</p>	<p>To derive EU-wide health based guidance values for BPA and other bisphenols (BPS and BPF). Find out how to feed this into an assessment of the TDI for BPA as set by EFSA. -&gt; WP5 / T 5.2, Finalise the establishment of HBM guidance values at European scale (WP5;Q: 8, 9)</p> <p>To relate exposure pathways including food pathway for BPA, BPS, BPF (possibly other BPs) to HBM data. -&gt; WP12</p> <p>To determine whether different regulatory controls across EU MS lead to different exposures.</p>
<p>9.</p> <p>Is it important to eliminate legacy BPA from material cycles (i.e. waste till receipt rolls) when implementing a circular economy in order to protect human health?</p>	<p>BPA (Cat. A)</p>	<p>As long term goal, it will be important to eliminate legacy BPA from material cycles (i.e. waste till receipt rolls) when implementing a circular economy in order to protect human health.</p> <p>Policy proposals should be developed to extend the focus to the whole group of bisphenols.</p>	<p>It is not clear whether these questions can be tackled within HBM4EU: Gather data on environmental persistence and the fate of bisphenols to determine exposure risks to humans and ecosystems. Studies that investigate photo-degradation and microbial degradation would provide understanding of environmental transformation products and fate of bisphenols. Correlate environmental monitoring data with HBM4EU data -&gt; WP12</p> <p>Use of available data (IPCHEM data) for improved exposure assessment and health impacts of bisphenols (WP5; Q: 8, 9)</p>

**Annex to table 9-2: Listing of available knowledge related to policy questions**

Substance	Available knowledge related to policy question
BPA	<p><b>Germany</b> 1995-2009,(20-29 yr.- urine &amp; plasma): [<b>Koch et al. (2012)</b>] &amp; 2003-2006 (GerES IV); (3-14 yr.- urine): [<b>Becker et al. (2009)</b>] (<i>rp</i>);</p> <p><b>Belgium, Flanders</b> 2007-2012 (14–15 yr.): [<b>Geens et al. (2014)</b>] (<i>rp</i>) &amp; 2011-2012 (DEMOCOPHES), (6-12yr., mothers, pregnant women- urine): [<b>Covaci et al. (2015)</b> and <b>3XG (2013)</b>] &amp; 2012-2015 (FLEHS 3), (50-65 yr.-urine): [<b>Steunpunt Milieu en Gezondheid (2015)</b>];</p> <p><b>Norway</b> 2012,(food &amp; estimated dietary exposure in adults): [<b>Sakhi et al. (2014)</b>];</p> <p><b>Greece</b> 2009-2011, (mother-child pairs: 2yr., pregnant women- urine): [<b>Myridakis et al. (Oct. 2015)</b>] &amp; 2011-2014, (children &lt;18yr., adults- hair): [<b>Tzatzarakis et al. (2015)</b>] &amp; 2012 (2.5-87 yr. X=49yr.- urine) [<b>Asimakopoulos et al. (Feb. 2014)</b>] &amp; 2014, (adult males, anonymous individuals- urine, serum)- analytical method: [<b>Myridakis et al. (Feb 2015)</b>, <b>Asimakopoulos and Thomaidis (2015)</b> and <b>Asimakopouloset al. (Jan. (2014)</b>] &amp; 2014 (Developing foetus, neonates, infants, children and adults- plasma, urine) -continuous lifetime model: [<b>Saringianis et al. (2014)</b>];</p> <p><b>Austria</b> 2008-2011, (mother- children pairs: 6-11yr., 25-50 yr.-urine): [<b>Hohenblum et al. (2012)</b>] &amp; 2010-2012, (6-15 yr., 18-64 yr., 65-79 yr.-urine): [<b>Hartmann et al. (2016)</b>];</p> <p><b>Sweden</b> 2008-2009 (food, young women-serum): [<b>Gyllenhammar et al. (2012)</b>] &amp; 2010-2011, (18-80 yr.-urine): [<b>Bjermo et al. (2013)</b>] &amp; 2010-2013, (17-19 yrs.-urine)-time series:[<b>Jönsson et al. (2014)</b>] &amp; 2011-2012 (DEMOCOPHES) (mother-child pairs: 6-11yr.,&lt;45yr.-urine): [<b>Larsson et al. (2014)</b>] 1996-2011, (first-time mothers-blood serum): [<b>Gyllenhammar et al (2012)</b> Tidstrend 1996-2011]</p> <p><b>Czech republic</b> 2015, (35.8±4.7 yr.-plasma, seminal plasma) analytical method: [<b>Vitku et al. (2015)</b>] &amp; 2000-2006, (canned foodstuffs, migration)-analytical method: [<b>Poustka et al. (2007)</b>] &amp; 1999-2000 (water samples &amp; river sediments): [<b>Stachel et al. (2003)</b>];</p> <p><b>France</b> 2011, 2013 (Blood, urine, amniotic liquid, tissue extracts) - analytical method: [<b>Lacroix et al. (2011)</b>, <b>Viguie et al. (2013)</b> and <b>Gayrard et al. (2013)</b>] &amp; 2013-2016, (Mother-premature infants-human breast milk): [<b>Deceuninck et al. (2015)</b>] &amp; 2003-2006, EDEN cohort (pregnant women-urine): [<b>Philippat et al. (2014)</b>] &amp; 2011 ELFE cohort (pregnant women on delivery-urine) [<b>Dereumeaux et al. (March 2016)</b> and <b>Dereumeaux et al. (Dec. 2016)</b>].</p> <p><b>France</b> 2014 -2016 Esteban (children &amp; adults, 6-74 yrs.- urine) [<b>Santé publique France, (Septembre 2019)</b>].</p> <p><b>Denmark</b> 2018 (children 8.5–16.1 yr.- urine) [<b>Carlsson et al. (2018)</b>] &amp; 2014 (young men- urine, blood &amp; semen) [<b>Lassen et al. (2014)</b>] &amp; 2006-2012 (children, adolescents, young men, and pregnant women- urine) [<b>Frederiksen et al. (2014)</b>] &amp; 2013 (rural &amp; urban mother-child pairs; urine) &amp; (children and adolescents; urine) [<b>Frederiksen et al. (2013)</b>]</p> <p><b>Spain</b> 2019 (adults, serum) subcohort of the Spanish European Prospective Investigation into Cancer and Nutrition (EPIC) [<b>Salamanca-Fernández et al. (2020)</b>]</p>
BPS	<p><b>Belgium, Flanders</b> 2012-2015 (FLEHS 3), (50-65 yr.-urine): [<b>Steunpunt Milieu en Gezondheid (2015)</b>] method development;</p> <p><b>Sweden</b> 1996-2011, (first-time mothers-blood serum): [<b>Gyllenhammar et al (2012)</b> Tidstrend 1996-2011];</p> <p><b>France</b> 2013-2016 (Mother-premature infants-human breast milk): [<b>Deceuninck et al. (2015)</b>];</p> <p><b>France</b> 2014 -2016 Esteban (children &amp; adults, 6-74 yrs.- urine) [<b>Santé publique France, (Septembre 2019)</b>].</p>

Substance	Available knowledge related to policy question
<b>BPF</b>	<p><b>Sweden</b> 1996-2011, (first-time mothers-blood serum): <b>[Gyllenhammar et al (2012) Tidstrend 1996-2011]</b>;</p> <p><b>Czech republic</b> 2000-2006, (canned foodstuffs, migration)-analytical method: <b>[Poustka et al. (2007)]</b> &amp; 1999-2000 (water samples &amp; river sediments): <b>[Stachel et al. (2003)]</b>;</p> <p><b>France</b> 2013-2016 (Mother-premature infants-human breast milk): <b>[Deceuninck et al. (2015)]</b>;</p> <p><b>France</b> 2014 -2016 Esteban (children &amp; adults, 6-74 yrs.- urine) <b>[Santé publique France, (Septembre 2019)]</b>.</p>
<b>BPB, BPAF, BPBP, BPC, BPC12, BPE, BPPH, BPM, BPP, BIS2, DHDPE, BPFL, BPZ, BP4,4'</b>	<p><b>France</b> 2013-2016 (Mother-premature infants-human breast milk): <b>[Deceuninck et al. (2015)]</b>;</p>

*rp = representative for the (respective) population*

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